

1 **Title** A Systematic Review of Approaches for Modelling Current and Future Impacts of Extreme  
2 Rainfall Events Using Green Infrastructure

3 Juliana Reu Junqueira <sup>a, \*</sup>, Silvia Serrao-Neumann <sup>a, b</sup>, Iain White <sup>a</sup>  
4 <sup>a</sup> Environmental Planning Programme, School of Social Sciences, University of Waikato, Hamilton, New Zealand  
5 <sup>b</sup> Cities Research Institute, Griffith University, Brisbane, Australia

6 **Abstract**

7 A range of modelling approaches has been developed to assess how green infrastructure  
8 could mitigate the effects of extreme rainfall events. This paper seeks to develop this agenda by  
9 reviewing how these modelling approaches incorporate, consider, and appraise information of  
10 value to land-use planning, policy, and practice to better understand why their implementation is  
11 infrequent and to help develop a research agenda. Our findings indicate that the information  
12 generated by current GI modelling approaches are not well integrated into the demands of land-  
13 use planning, and may more reflect the information availability than useability. We find that  
14 modelling outputs do not tend to generate the type of high resolution information covering the  
15 appropriate spatial and temporal scales that is needed to best support planning decisions.  
16 There are also gaps in the assessment of future climate risks, such as increased rainfall  
17 intensity, and how this links to future pressures, such as escalating urban growth and  
18 development demands, and the interaction between the two areas. The paper concludes that to  
19 increase the implementation of green infrastructure, modelling researchers should work more  
20 closely with decision-makers to better link data on the effects of GI to the politics involved in  
21 their implementation in planning decisions, particularly how trade-offs occur over different scales  
22 and times, and between sectors.

23 **Key words:** flooding, low-impact development - LID, stormwater management, land-use  
24 planning, urban drainage, climate change

25

26 **Introduction**

27 Flood damages have increased through time and flood risks continue to rise due to  
28 issues such as ongoing urbanisation in flood-prone areas, a lack of integration between land-  
29 use planning and flood risk management, and ageing stormwater infrastructure unable to deal  
30 with intensified runoff loads (Sohn et al., 2019). Recently, green infrastructure<sup>1</sup> (GI) has

---

<sup>1</sup> Green infrastructure refers to a multifunctional and interconnected network of natural areas and open

31 emerged as a promising flood risk management alternative to grey infrastructure<sup>2</sup>, particularly  
32 given its potential to deal with extreme rainfall events and associated flood risks (Arnone et al.,  
33 2018). It is argued that GI can minimise flood risks because it may reduce demand loads on  
34 stormwater infrastructure, especially GI alternatives that can increase rainfall infiltration and  
35 retain and store stormwater runoff (Jia et al., 2015).

36 Nonetheless, widespread GI implementation still faces resistance. As an emerging  
37 approach, there is limited empirical evidence of its effectiveness in managing flood risks (Thorne  
38 et al., 2018). In particular, the selection of GI alternatives requires significant experimental and  
39 modelling investigations to test their efficiency in addressing flood risks, which are context  
40 specific and not easily transferable (Baek et al., 2015). There are also unresolved questions  
41 related to both implementation and maintenance costs (Jia et al., 2015). Given this context, it is  
42 unsurprising there is limited adoption and implementation of GI alternatives into land-use  
43 planning (O'Donnell et al., 2017; Pappalardo et al., 2017). The deficit between potential and  
44 practice suggests that there is a substantial knowledge gap, not only relating to GI  
45 effectiveness, but also in promoting and prioritising GI in long-term growth plans, planning  
46 policies, and decision-making processes. This is becoming ever more urgent considering the  
47 growing threat from rainfall intensification as a result of climate change (Gill et al., 2007; Hislop  
48 et al., 2019), and the new urbanisation demands associated with addressing the housing crisis  
49 in many countries (White and Nandedkar, 2019). This also suggests that to better inform land-  
50 use plan-making and plan-implementation, information on the effectiveness of GI in flood  
51 minimisation needs to provide context specific granularity at both spatial and temporal scales,  
52 along with more practical issues, such as relating to their space requirements, to gain support  
53 from developers and decision-makers.

54 The spatial granularity is important because such information needs to be translated into  
55 the various land-use overlays and risk assessments routinely used in development control and  
56 land-use planning (LeGates et al., 2009; Zevenbergen et al., 2008), as well as to clarify potential  
57 trade-offs between policy arenas (e.g. housing, nature conservation, transport) (Busscher et al.,

---

spaces managed by humans that aims to preserve principles and functions of natural ecosystems and related ecosystem services (Benedict, M.A., McMahon, E.T., Mark, A.t.C.F., Bergen, L., 2006. Green infrastructure : Linking landscapes and communities. Washington: Island Press, Washington, Salata, K., Yiannakou, A., 2016. Green infrastructure and climate change adaptation. Tema 9(1), 10-27.).

<sup>2</sup> Grey infrastructure relates to human-made infrastructure and includes measures relating to build structure such as dams and levees (Soz, S.A., Kryspin-Watson, J., Stanton-Geddes, Z., 2016. The role of green infrastructure solutions in urban flood risk management. Urban Floods Community of Practice.).

58 2019). The temporal granularity is required because this information needs to be future-oriented  
59 in order to balance costs between current and future ratepayers, and consider the range of  
60 climate change futures to help development avoid, reduce or better manage flood risks  
61 (Campbell, 2006; Quay, 2010; Woodward et al., 2014). While it seems obvious that strategic  
62 land-use planning<sup>3</sup> considers future impacts and risks, traditionally, decisions have been made  
63 based on past events (Duinker and Greig, 2007; Kelly et al., 2004; O' Brien and O' Keefe,  
64 2013). For example, there is a tendency to see a 'Tyranny of the Present' (White and Haughton,  
65 2017) as many localities use historical flood events to establish set rules and regulations for  
66 development control, and struggle to incorporate uncertain future risks into long-term growth  
67 plans (Schuch et al., 2017). Lastly, when applied to urban form and decision-making, the  
68 analysis of GI should also assess their cost-effectiveness (Lemes de Oliveira, 2019). This  
69 information is particularly useful to assess not only how GI implementation can be more  
70 economically efficient in comparison to other measures, but also whether the adoption of GI  
71 alternatives is a realistic and politically acceptable use of that particular space (Block and  
72 Strzepek, 2010).

73 As GI became an emerging field of study, a plethora of modelling approaches has been  
74 developed to assess its effectiveness, including to minimise flood risks (Alves et al., 2019). A  
75 consistent analytical framework for assessing the effectiveness of GI in planning, however, is  
76 still missing, perhaps due to the variety of functions GI can serve and the different user groups  
77 (Salata and Yiannakou, 2016). This paper seeks to address this gap in knowledge by providing  
78 a systematic literature review incorporating (i) how different approaches for modelling current  
79 and future impacts of rainfall events incorporate GI, (ii) how models addressed differing spatial  
80 and temporal scales in their analyses, and (iii) what aspects related to GI design and  
81 assessment were considered by differing models. In doing so, we also sketch out a potential  
82 research agenda; one which we hope holds potential to help close the gap between both  
83 modelling and planning disciplines, and between science and implementation more generally.

## 84 **Method**

85 Given the range of studies and methodologies aiming to model the current and future  
86 impacts of rainfall events using GI alternatives, we undertook a systematic review of literature to

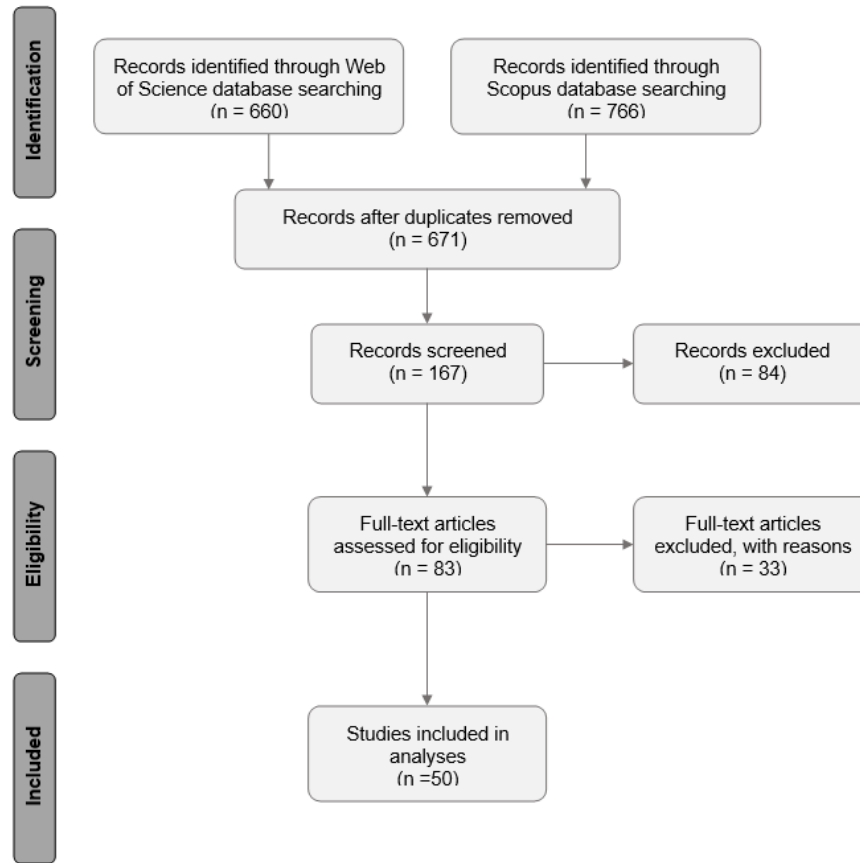
---

<sup>3</sup> Strategic land-use planning is a systematic process that aims to prepare cities for future social and environmental change and ensure current and future development and growth management is planned accordingly (Khalil, H.A.E.E., 2012. Enhancing quality of life through strategic urban planning. Sustainable cities and society 5, 77-86.).

87 assess the parameters used by different modelling approaches (Paré et al., 2015). A systematic  
88 literature review is an established method for finding, analysing and evaluating information on a  
89 specific topic (Bilotta et al., 2014). While it was first applied to the medical sciences, the method  
90 has gained popularity in studies related to the environment (Mallett et al., 2012). In particular, it  
91 is now accepted as an effective method to summarise information related to a specific field  
92 because it provides textual evidence, helps remove researcher's bias, and adheres to scientific  
93 rigour, especially reproducibility, transparency and reliability principles (Bilotta et al., 2014;  
94 Mallett et al., 2012). This paper followed the steps proposed by Bilotta et al. (2014): pre-  
95 identification, identification of potential studies, screening and eligibility, and inclusion.

## 96 **Pre-identification**

97 A search of literature first investigated studies that incorporated GI in flood modelling to  
98 assess differing approaches for modelling current and future impacts of extreme rainfall events.  
99 These were identified through a systematic search of peer-reviewed literature (Web of Science  
100 and Scopus databases) following the protocol of Preferred Reporting Items for Systematic  
101 Literature Reviews and Meta-Analyses (PRISMA) (Moher et al., 2009). The initial search terms  
102 used were: (i) "green infrastructure" + "modelling", and (ii) "LID" + "modelling". Depending on the  
103 central goal of the study, a wide range of approaches to flood modelling were found: from vast  
104 and sophisticated mathematical and statistical models, through to multi-methods models  
105 combining numerical and non-numerical evaluations. This first screening resulted in more than  
106 5,000 returns. It was, therefore, necessary to narrow that down to those which considered  
107 modelling impacts of extreme rainfall events using GI options. A second search was therefore  
108 undertaken to identify key definitions of terms and concepts relating to modelling impacts of  
109 extreme rainfall events using GI techniques. A series of steps then followed to refine the initial  
110 search results and select studies for further content investigation (see Figure 1).



111

112 *Figure 1 - Flow diagram showing the different steps of the systematic literature review refinement process*

113

### 114 **Identification of potential studies**

115 The pre-identification step informed the selection of key terms used to refine the search  
 116 and narrow down initial results to a manageable size. Terms were now defined based on three  
 117 questions:

- 118 (i) How do different approaches for modelling current and future impacts of rainfall  
 119 events incorporate GI?
- 120 (ii) How do models address differing spatial and temporal scales?
- 121 (iii) What aspects related to GI design and assessment are being considered?

122 The suite of terms used included a combination of relevant terms, namely: “modelling” +  
 123 “Green Infrastructure” + “storm” or “modelling” + “LID” + “storm” or “modelling” + “WUSM” +  
 124 “storm” or “modelling” + “source control” + “storm” or “impacts” + rainfall” + “Green  
 125 Infrastructure” or “impacts” + “rainfall” + “LID” or “impacts” + “rainfall” + “WUSD” or “impacts” +

126 “rainfall” + “source control” or “Green Infrastructure” + “flood risk” + “management” or “LID” +  
 127 “flood risk” + “management” or “WUSD” + “flood risk” + “management” or “source control” +  
 128 “flood risk” + “management”. The terms relating to GI aspects followed the most used  
 129 terminologies reported in the study by Fletcher et al. (2015), including: Low Impact Development  
 130 (LID), Green Infrastructure, Source Control, and Water Sensitive Urban Design (WSUD). Best  
 131 Management Practices (BMPs) were excluded because their primary focus is on pollution  
 132 prevention (Fletcher et al., 2015) and not flood risks or effects from extreme rainfall events. The  
 133 collocation of keywords followed a sequence guided by key terms related to the key concepts  
 134 underpinning this paper: flood modelling, climate change impacts, and GI terminologies. The  
 135 clear identification of such terms ensures the method is reproducible and the outcomes  
 136 transparent. This yielded 660 results in the Web of Science and 766 in Scopus published  
 137 between 2013 and 2019. A summary of the search results based on different combinations of  
 138 search terms is shown in Table 1.

139 *Table 1 - List of databases, keywords, and number of results used in the preliminary search.*

<b>Search Terms</b>	<b>Web of Science</b>	<b>Scopus</b>
modelling AND green infrastructure AND storm	85	130
modelling AND LID AND storm	167	172
modelling AND WSUD AND storm	7	11
modelling AND source control AND storm	26	39
impacts AND rainfall AND green infrastructure	79	66
impacts AND rainfall AND LID	170	205
impacts AND rainfall AND WSUD	12	10
impacts AND rainfall AND source control	18	21
green infrastructure AND flood risk AND management	78	85
LID AND flood risk AND management	13	22
WSUD AND flood risk AND management	3	5
source control AND flood risk AND management	2	3
<b>Total</b>	<b>660</b>	<b>766</b>

140

141 **Screening and eligibility**

142 When combined, and after removal of duplicates, the dataset included 671 publications  
143 from different fields. These included environmental management, hydrology, urban ecology,  
144 industrial ecology, ecological engineering, flood risk management, environmental modelling, and  
145 ecological modelling. Titles and abstracts of the 671 publications were assessed to determine  
146 whether they were suitable for inclusion in the core review. This enabled the identification of  
147 studies investigating the use of single and/or combined GI alternatives as a way of minimising  
148 flood impacts. This step resulted in 167 publications which were then screened in full. Those  
149 with no primary focus on flooding (e.g. water quality, material performance, numerical modelling,  
150 groundwater, runoff coefficients, health, rain gage data, soil, evapotranspiration, erosion, or  
151 thermal characteristics) were excluded, resulting in the selection of 50 studies to be included in  
152 the final review (see Supplementary Material A for a full list).

153 **Content analysis**

154 The next stage was a content analysis. This approach is designed to understand the  
155 essence of written or visual sources by systematically assigning their content to pre-defined,  
156 comprehensive categories, and then both quantifying and interpreting the outcomes (Byrne, 2017;  
157 Neuendorf, 2017; Payne and Payne, 2004). Using NVivo 12 Pro, the content of selected articles  
158 (n=50) was first analysed and grouped based on the aim of the studies (see Table 2), as the  
159 objective of the study can be a determinant on the modelling approach applied.

160 *Table 2 - Key aims identified across publications selected*

<b>Aim of study</b>	Calculation and simulation of the hydrological performance of GI
	Application of varying types and combinations of GI
	Estimating flood risk/damage
	Governance perspective of the implementation of GI
	Application of GI for Urban Water Management (UWM)

161

162 The content analysis then followed a coding frame (Byrne, 2017) based on the set of  
163 exploratory questions. The exploratory questions intended to investigate the extent to which the  
164 modelling approaches generated outputs that could inform land-use planning decisions. To this  
165 end, the questions provided an analytical framework to investigate how publications dealt with  
166 aspects related to spatial and temporal scales, as well as GI design and assessment in their  
167 modelling approaches (see Table 3). This was designed to facilitate the link to land-use planning

168 processes and practices, and to identify similarities and differences between articles. They also  
 169 enabled the assessment of types of data sets used by publications to gain a more advanced  
 170 understanding of how each model dealt with current and future impacts of rainfall events, and  
 171 how this was applied to different GI types and combinations. The coding frame was regularly  
 172 reviewed by the authors based on the actual content of the publications to improve its reliability  
 173 and content coverage so as to ensure all evidence was collated and summarised (see  
 174 Supplementary Material B for further information regarding the coding process).

175 *Table 3 - Themes and exploratory questions used to analyse publications content*

Theme	Exploratory questions
<b>Spatial Scale</b>	What is the spatial scale covered by the study (e.g. whole catchment, single waterway)?
	What kind of urbanisation pattern does the study focus on (e.g. urban infill, peri-urban, or greenfield development)?
	What types of land-use are included in the study (e.g. residential, commercial, industrial)?
<b>Temporal Scale</b>	What range of historical rainfall data is used by the study?
	What range of future rainfall data projection is used by the study (e.g. climate change projections)?
	How does the study deal with past and future land-use/land cover change?
<b>GI Design and Assessment</b>	How does the study design and assess GI alternatives (e.g. single option, multiple options not combined, or combined multiple options)?
	How does the study deal with data paucity and quality?
	What GI implementation aspects are discussed (e.g. cost analysis, an optimal combination of GI, runoff management, suitable locations)?

176

## 177 **Results**

178 The global reach of the selected publications shows the fundamental challenge rainfall  
 179 extremes present for urban areas, with a predominance of studies carried out in China (19) and  
 180 the U.S.A. (13). Few publications (three out of 45) focused on more than one case study  
 181 location, including across multiple countries. As may be expected in an emergent field, most of  
 182 the studies were published between 2017 and 2019 (36 out of 50), with 2019 accounting for  
 183 seven, 2018 eighteen, and 2017 eleven. Detailed information about all 50 studies, including key  
 184 findings related to the application of the exploratory questions to selected publications and  
 185 overall results from the content analysis can be found in the Supplementary Material C, D, E  
 186 and F.

### 187 **Choice of modelling approach based on study aims**

188 While GI has a broad remit, two out of the five aims identified in Table 2 were  
 189 predominant amongst the publications. These focused on providing data relating to the



190 calculation and simulation of the hydrological performance of GI, and the application of varying  
191 types and combinations of GI. The objective of the study was found to be a crucial component in  
192 the choice of the approach used to model extreme rainfall events. In particular, two main  
193 families of approaches were identified for modelling impacts of rainfall events: hydrological  
194 models and social-technical perspectives.

### 195 **Types of hydrological models and social-technical perspectives**

196 While varied in the way GI was incorporated, hydrological models dominated the  
197 strategies used to evaluate the effects of rainfall events in comparison to social-technical  
198 perspectives (45 out of the 50 studies used hydrological models as their key modelling  
199 approach). All of the publications using hydrological models focused on two aims, namely:  
200 calculation and simulation of the hydrological performance of GI, and application of varying  
201 types and combinations of GI. For example, Cipolla et al. (2016) used a hydrological model to  
202 analyse the performance of green roofs as an opportunity to manage stormwater runoff after  
203 urbanisation, simulating a long-term hydrologic response (over one year), and comparing the  
204 results to an adjacent impervious roof of the same scale. Goncalves et al. (2018) assessed to  
205 what extent different GI alternatives were able to reduce flood risk by developing feasible  
206 scenarios where GI units were placed throughout the study area and comparing the results  
207 against a baseline scenario. The approach given by the publications using hydrological models  
208 connect the use of GI options to the hydrology and engineering fields, rather than to land-use  
209 planning or as an instrument to minimise the effects of climate change.

210 Hydrological models are effective in describing water processes at the catchment or  
211 waterway level (e.g. precipitation, flood, runoff) and in determining the requirements for the  
212 conservation and management of water resources (Parra et al., 2018; Rusli et al., 2015). They  
213 are usually designed for multiple spatial and temporal scales and allow a variety of GI design and  
214 assessment options. Potentially, this is highly relevant to inform land-use planning decisions as  
215 they can provide various scenarios to evaluate different scales and GI assessments. However,  
216 they tend to oversimplify the real world (Dietrich et al., 2016) and modelled processes (Parra et  
217 al., 2018), as well as ignore the exchange of groundwater (Pellicer-Martínez et al., 2015). They  
218 also tend to emerge from technical disciplines that may not typically capture and consider the  
219 political trade-offs that characterise discussions over differing land-use futures.

220 Three studies adopted social-technical perspectives with specific aims (e.g. estimating  
221 flood risk/damage, governance perspective of the implementation of GI, and application of GI for

222 Urban Water Management). Liu and Jensen (2018) used a multi-level perspective to investigate  
223 the role of GI in urban water management practices of five cities famous for their progressive  
224 approach: Singapore, Berlin, Melbourne, Philadelphia, and Sino-Singapore Tianjin Eco-city in  
225 China. The authors adopted an analytical framework based on transition theory to analyse the  
226 main challenges for GI implementation. Schifman et al. (2017) suggested a Framework for  
227 Adaptive Socio-Hydrology in which GI planning and implementation advanced from a purely  
228 hydrological viewpoint to an integrated social-hydrological approach. In contrast to other  
229 research, their study sought to provide an integrated, multifaceted decision-making process  
230 rather than focusing on highly centralised stormwater management aspects that are often  
231 disconnected from other elements of urban landscapes. O'Sullivan et al. (2015) used Life Cycle  
232 Assessment (LCA) to quantify the environmental impacts associated with materials,  
233 construction, transport and maintenance of GI alternatives. The authors evaluated the LCA of GI  
234 alternatives to provide a guiding tool for practitioners such as planners, engineers and decision-  
235 makers. The analysis of these three publications indicated that the adoption of social-technical  
236 perspectives can cover several aspects, but they did not include issues connected to why GI  
237 has had little practical influence as a means to mitigate extreme rainfall events.

238           There is potential in this area though. Social-technical perspectives are efficient in  
239 analysing elements dependent on the behaviour of those who use it (Lamond and Everett,  
240 2019). Therefore, these approaches can provide a deeper understanding of how GI design  
241 alternatives, co-creation and incorporation into dynamic socio-ecological-technical processes  
242 can be improved (Ward et al., 2019) and how GI alternatives impact (or are impacted by) land-  
243 use planning. Such strategies can enhance the awareness of practitioners, individuals and  
244 communities, which is important in recognising a diversification of assessments (Ward et al.,  
245 2019). In particular, they are useful in estimating flood risk/damage, evaluating governance  
246 arrangements for implementing GI, and the application of GI for urban water management.  
247 Some weaknesses to note include not accounting for co-benefits produced (O'Sullivan et al.,  
248 2015), missing wider perspectives or data that could have linked GI design with existing  
249 resources (Schifman et al., 2017), and challenges in incorporating future rainfall data and cross-  
250 scale dynamics.

251           Two studies used a combination of approaches (hydrological modelling and social-  
252 technical perspectives). Song and Chung (2017) developed a multi-criteria decision analysis  
253 framework in order to prioritise sites and types of GI alternatives, coupling it with a hydrological  
254 model called Storm Water Management Model (SWMM). While Yang and Chui (2018) first ran a

255 hydrological simulation using SWMM and then adopted a relative performance evaluation  
256 method (RPE) to assess hydro-environmental impacts of GI techniques in a case study in  
257 Brooklyn, New York. While rare, studies aimed at calculating and simulating the hydrological  
258 performance of GI, or applying various types and combinations of GI, can bring benefits from  
259 the use of both families of approaches (Goncalves et al., 2018; Schubert et al., 2017; Wang et  
260 al., 2019).

261 A range of models was used by hydrological and social-technical approaches to assess  
262 current and future impacts of rainfall events, but there was a clear dominance in the use of  
263 SWMM. Some studies used single models (e.g. Ahiablame and Shakya (2016) and Eckart et al.  
264 (2018) focused only on SWMM, and Jia et al. (2015) utilised SUSTAIN), while others utilised a  
265 set of models. For example, Zhu and Chen (2017) used SWMM and a particle swarm  
266 optimization algorithm. Song and Chung (2017) used a combination of models, including  
267 SWMM, TOPSIS, Delphi method, and multi-criteria decision analysis (MCDA). By using a  
268 combination of models or utilising different complementary models it is possible to minimise the  
269 gaps present in each of the models, and linking them to land-use planning demands and  
270 parameters, resulting in a more realistic simulation.

## 271 **Application of hydrological models**

### 272 ***Spatial scale***

273 The majority of studies using hydrological models (35 out of 45) assessed waterways at  
274 different spatial scales, typically single waterways (26) or whole catchments (9). For example,  
275 Yau et al. (2017) based their research on Waterway Ridges, a four-hectare pilot urban  
276 development project in Singapore with a single waterway (Punggol Creek). Their study  
277 investigated different types of GI alternatives which were innovatively integrated into the design  
278 at the precinct level. With respect to whole catchments, Ahiablame and Shakya (2016) carried  
279 out a study in the City of Normal-Sugar Creek Watershed, McLean County in Central Illinois,  
280 due to its high urbanisation rate and good data availability. Their study assessed to what extent  
281 large scale adoption of GI alternatives in an urban watershed could increase flood reduction  
282 capabilities.

283 There was variation in the way the different models used by the publications dealt with  
284 the spatial scale, but the majority of the studies used SWMM (25 out of 37). Damodaram and  
285 Zechman (2013) used a combination of links and nodes representing the stormwater

286 infrastructure, composed of a box and circular storm sewers, open channels, and a single  
287 natural waterway in order to design a GI scenario for managing peak flow alterations. Two  
288 publications used the L-THIA-LID model to study the entire catchment. For example, Eaton  
289 (2018) used various combinations of GI strategies to evaluate stormwater reduction for the  
290 whole Alley Creek watershed in New York City.

291           Concerning urbanisation, 16 out of 45 publications studied waterways within urban infill  
292 developments. For example, Baek et al. (2015) studied a commercial area in Gwangju, Korea,  
293 characterised by high imperviousness (approximately 85%) to define an optimal combination of  
294 GI alternatives to reduce flood and improve water quality. Nine out of 45 publications  
295 investigated entire catchments located in peri-urban areas. Kong et al. (2017) used a case study  
296 located west of downtown Bazhong which was predominantly rural, covered by farmland  
297 (49.2%), forest (42%), with housing, roads, and water bodies only representing 3% of the total  
298 area (838 ha). Their research focused on several sub-catchments (80 sub-catchments in the  
299 pre-development state and 118 sub-catchments in the urban development scheme) to model  
300 the potential effects of large-scale implementation of GI alternatives. Few studies assessed  
301 waterways in greenfield developments (four out of 45 studies). None used a single waterway as  
302 their spatial scale. For example, Bai et al.'s (2018) study area was Sucheng District, located in  
303 the north of Jiangsu Province, China, which was divided into 83 sub-catchments and the river  
304 channel (treated in the study as a pipeline) to evaluate the impact of GI strategies on the  
305 reduction of surface runoff and flood volume.

306           Urbanisation patterns appear not to be a key determinant employed by the studies in the  
307 selection of models used to assess GI strategies; rather their selection relied on models'  
308 capability and richness of functions with SWMM being the predominant choice. Some of the  
309 main capabilities included the integration with GIS, the number of GI alternatives available, and  
310 the ability to input rainfall data. However, none of the studies included urbanisation patterns.  
311 Interestingly, while models do not present specific capabilities or functions concerning  
312 urbanisation patterns, these were used as external information to determine the size of the  
313 study area and the selection of GI alternatives. For example, Liao et al. (2013) used SWMM to  
314 study a highly urbanised area in Shanghai, China, to simulate runoff reduction, peak flow rate  
315 reduction and waterlogging volume reduction using five GI alternatives. Kong et al. (2017) used  
316 SWMM to study a peri-urban area with predominantly rural cover in the west of downtown  
317 Bazhong to assess the hydrological responses of stormwater runoff related to four different  
318 land-use conversion scenarios using GI alternatives. Burszta-Adamiak and Mrowiec (2013)

319 applied SWMM to an experimental greenfield development site in Poland to investigate the  
320 hydrological performance of green roofs in reducing surface runoff and flood risks due to  
321 snowmelt and heavy rainfall.

322 Land-use types were categorised and investigated by the majority of the studies (28 out  
323 of 45). These predominantly included residential (19), commercial (10), industrial (8) and open  
324 space (10). Most of the studies (25 out of 28) used mixed land-use types. Li et al. (2019)  
325 focused on residential areas with condominiums and parking lots, residential areas with  
326 condominiums but no parking lots, residential areas with single-family houses, industrial areas,  
327 and commercial areas for testing GI options to assess the impacts of land-use on runoff.  
328 Residential land-use was evaluated by three studies as a single land-use type. For example,  
329 Zhu and Chen (2017) used a typical highly-developed residential area in Guangzhou, China, to  
330 evaluate the effects of GI techniques on urban flooding under different rainfall intensities.

### 331 ***Temporal scale***

332 Most of the studies using hydrological models assessed rainfall at different temporal  
333 scales (36 out of 45) - that is, historical (28), historical and future (five) or future rainfall data  
334 (two). The majority of the studies considering historical rainfall data in their models used up to  
335 one decade as their temporal analyses (19 out of 28), followed by three decades or more  
336 (seven), and two decades (two). Using historical data, Palla and Gnecco (2015) calibrated and  
337 validated their model based on seven rainfall events collected between February and May 2005  
338 to analyse how GI techniques performed as source control for peak flow reduction, volume  
339 reduction, and hydrograph delay. Utilising historical and future data, Zhang et al. (2019) applied  
340 a downscaling method to assess the reliability of GI techniques in pollution reduction, flow  
341 frequency mitigation, and potential to provide an alternative water supply comparing various  
342 future climate conditions (2040-2049) against the base-line period (1995-2004). Using future  
343 projections, Zhang et al. (2018) assessed hydrological effects and performance of GI options in  
344 four urban catchments based on differing scenarios for the period between 2040 and 2059.

345 Some studies (10 out of 45) used predefined rainfall models in their simulation without  
346 considering any temporal scale. For example, Bai et al. (2018) adopted the Chicago Rainfall  
347 Model, a widely used uneven rainfall model based on the intensity-duration-frequency  
348 relationship, to simulate the effect of four different types of GI scenarios on urban flooding.

349 Climate change projections were specifically assessed by few studies (only seven out of  
350 36). Five considered historical and future rainfall data, and two focused only on future rainfall  
351 projections. Concerning future rainfall, studies focused on ranges of three or more decades  
352 (seven). Chen et al. (2017) considered climate change conditions for the 2021-2040 and 2061-  
353 2080 scenarios to identify optimal GI layout and enable decision-making to minimising flood risk.  
354 Some studies (three out of seven) used both historical and future data. Wang et al. (2019) used  
355 selected climate change scenarios for 2020-2050 and compared them with scenarios from  
356 1997-2000 to assess the performance of porous pavement and bioretention cells for stormwater  
357 management.

358 Few studies (12 out of 45) assessed the temporal scale in association with potential land  
359 use or land cover change. Mao et al. (2017) reclassified land use data to match time-series  
360 input (2008) in order to assess the ecological benefits of aggregated GI techniques for  
361 stormwater runoff control. Li et al. (2019) tested two land-use scenarios simulations (2011 and  
362 2050) to assess the effects on surface runoff of GI techniques under different land uses.

### 363 ***GI design and assessment***

364 There was variation in the number of GI options considered by the publications, with  
365 most studies incorporating three or more GI techniques into their models (25 out of 45). The  
366 alternatives most investigated were bioretention (30), permeable pavement (29), green roofs  
367 (24), and vegetated swales (15). Wang et al. (2016) simulated future scenarios to assess cost-  
368 effectiveness of bioretention on stormwater as a response to urbanisation and climate change.  
369 Although using a generic methodology, the authors acknowledged that the performance of  
370 bioretention in an urban catchment is more efficient for dealing with urbanisation changes than  
371 for climate change effects. Wang et al. (2019) assessed the performance of bioretention cells  
372 and permeable pavements for stormwater management as a response to climate change in  
373 Guangzhou, China. The authors found that both permeable pavement and bioretention cells  
374 could reduce runoff volume and peak discharge in response to rainfall events in short periods,  
375 but not for heavy storms with a longer return period. Bai et al. (2018) tested four different types  
376 of GI scenarios to control runoff in urban areas: (i) no GI technique; (ii) GI technique based on  
377 infiltration; (iii) GI technique based on water storage; and, (iv) GI technique based on a  
378 combination of infiltration and water storage. The combined model (infiltration + storage)  
379 presented the best performance in runoff reduction. The research also acknowledged that GI

380 alternatives could reduce the risk of urban flood impacts caused by extreme rainfall events - if  
381 planned correctly.

382           The difference between the number of GI alternatives used by the hydrological models  
383 and the socio-technical perspective seems to rely more on authors' choice than on some  
384 features of the models. Although there are some pre-defined GI options available for  
385 hydrological models, none of the publications have used all the possible alternatives in the  
386 same study. In practice, differing GI features do, however, take up very different space within a  
387 city, and so there are considerable political and resource implications of these modelling  
388 selections.

389           The majority of studies assessed combined GI options (31 out of 45). Li et al. (2018)  
390 selected GI alternatives based on differing characteristics of the study area, including buildings  
391 with flat green roofs, buildings with sloping roofs and rain barrels; parking lots and impervious  
392 roads with porous pavement; and existing green gardens along roadways converted to  
393 bioretention cells. Zhu and Chen (2017) investigated the effects of rain gardens and bioretention  
394 cells on flood control under different rainfall scenarios. These included different return periods  
395 (one year, five years and ten years), rainfall durations (1h, 1.5h and 2h) and rainfall peak  
396 coefficients (0.375, 0.5 and 0.8) to evaluate the changes before and after GI. The outcomes  
397 demonstrated that with the increase in rainfall peak coefficient, intensity and duration, the  
398 control effects of both GI options tend to be reduced.

399           Fewer studies acknowledged single options of GI (eight out of 45) or multiple but not  
400 combined GI alternatives (three out of 45). Ercolani et al. (2018) analysed the implementation of  
401 green roofs as source control solution for mitigating the impacts of urbanisation. The authors  
402 assessed the effect of four spatially homogeneous installations of green roofs (25%, 50%, 75%,  
403 and 100% of roofs area covered) and a spatially heterogeneous targeted conversion  
404 (concentrating green roofs where conduits were more prone to filling) using six storm events  
405 differing in both duration and return period. The heterogeneous scenarios presented better  
406 results in terms of reduction rates of peak flow and volume at the network outlet. Li et al. (2019)  
407 explored the cost-efficiency scenario for runoff water quantity reduction of green roofs, rain  
408 cisterns, rain barrels, porous pavement, bioretention cells, grassed swales, wet ponds, and dry  
409 ponds. The authors tested the efficiencies of combined individual GI types and found that the  
410 combination of grassed swales, rain barrels, dry ponds, and porous pavement were the most  
411 cost-efficient scenario for reducing the amount of water runoff.

412 Data input and calibration is a complex step in GI design and assessment, and several  
413 studies used different levels of simplification to calibrate and run their models (23 out of 45). Hu  
414 et al. (2019) tested a range of storm events with differing intensity-duration-frequency curves for  
415 rainfall to evaluate the hydrological performance of GI techniques using an empirical formula of  
416 rainfall and Chicago hyetograph method for rainstorm design. No effort was made to calibrate  
417 the model due to the lack of observed data. Furthermore, model parameter values were  
418 obtained from the published literature. Fu et al. (2018) used remote sensing to analyse land-use  
419 and land cover change types and runoff variations in response to urbanisation at different  
420 spatial scales using scenarios with and without GI alternatives, including basin, watershed, and  
421 city scales. Their results demonstrated that at the basin scale, land-cover changes interpreted  
422 from satellite images were very helpful for identifying watersheds with urbanisation hotspots that  
423 might have larger runoff outputs. However, at the watershed scale, the resolution of the land  
424 cover data was too low and needed to be replaced with observed land-use data using  
425 sophisticated hydrological modelling to evaluate runoff for scenarios with and without GI  
426 alternatives at different spatial scales.

427 The most recurrent GI implementation aspects investigated by the studies were: (i) cost  
428 analysis (14 out of 45); (ii) optimal combination of GI alternatives (10 out of 45); (iii) runoff  
429 management efficiency (43 out of 45); and, (iv) suitable locations for implementing GI  
430 alternatives (four out of 45). Eckart et al. (2018) used three storm events (with different return  
431 periods) to test five different GI implementation scenarios, including rain barrels, porous  
432 pavement, bioretention options (both engineered bioretention and simple rain gardens), and  
433 infiltration trenches. After evaluating the performance of GI for all three design storms for each  
434 of the five GI implementation scenarios, the authors concluded that infiltration trenches would  
435 be the most cost-effective GI alternative for reducing peak flow. Baek et al. (2015) proposed the  
436 optimal size for each GI alternative (bioretention, green roof, infiltration trench, porous  
437 pavement, rain barrel, and vegetative swale) by conducting intensive stormwater monitoring and  
438 numerical modelling in a commercial area in Korea for minimising flood effects.

439 With respect to runoff management efficiency, Sun et al. (2014) showed that GI  
440 techniques were effective in controlling stormwater flow for small rainfall events but not for  
441 larger rainfall events. Qin et al. (2013) analysed the effects of three GI alternatives (permeable  
442 pavements, green roofs and swales) to control urban flooding and compared the results with  
443 conventional drainage system design. Their results indicated that the tested GI scenarios were  
444 more effective in flood reduction during heavier and shorter storm events, and that permeable



445 pavements performed best during a storm event with a middle flow peak, green roofs performed  
446 best with a late peak, and swales performed best with an early peak. Suitable locations for GI  
447 alternatives were briefly assessed (four out of 45). Eaton (2018) identified that the search for the  
448 optimally effective GI combination should start with the most effective techniques for each land-  
449 use type. Results were responsive to the relative location of various land-use types and showed  
450 that bioretention and rain gardens provided the most significant reduction on a residential  
451 watershed. A key observation is that the low prevalence of location and combination aspects  
452 within these models, despite being a core consideration for land-use planning, may help explain  
453 why these elements have struggled to be implemented in practice.

#### 454 **Application of the social-technical perspective**

##### 455 ***Spatial scale***

456 Spatial scale was assessed by one out of five studies applying social-technical  
457 perspectives. Song and Chung (2017) used a case study at a university campus in Seoul, with a  
458 single waterway to test a multi-criteria decision analysis (MCDA) framework in order to prioritise  
459 sites and types of GI alternatives.

460 Regarding urbanisation patterns, two out of five publications focused on urban infill.  
461 Schifman et al. (2017) proposed a decision-making framework to allow a more interactive  
462 process for the installation of GI rather than a 'one-size-fits-all' approach by comparing  
463 Cleveland and Atlanta, USA. The study showed that a multi-stakeholder, integrated,  
464 decentralised network with co-decision project plan resulted in enhanced multifunctionality,  
465 enabling resilience in urban systems at multiple scales. Song and Chung (2017), using a case  
466 study in South Korea, tested a system capable of simulating and ranking multiple GI scenarios  
467 based on hydrological aspects along with social factors, using scenario performance values.

468 Land-use type is not well represented in social-technical studies. This was only  
469 mentioned by Song and Chung (2017), who briefly stated 92.7% of the study area was covered  
470 by building, roads, and green spaces, without linking this discussion to long term plans,  
471 development pressures, or political priorities. Again, this may help explain why modelling results  
472 are struggling to influence land-use planning decisions.

##### 473 ***Temporal scale***

474 Fewer studies using social-technical perspectives assessed rainfall at different temporal  
475 scales (two out of five), that is, up to one decade of rainfall data (one out of five) and three or

476 more decades (one out of five). None of the publications assessed climate change projections  
477 relating to future rainfall data. Yang and Chui (2018) used the hourly precipitation record from  
478 1969 to 2013 at JFK International Airport in New York City) to simulate a long-term hydrological  
479 performance of various GI techniques before applying a relative performance evaluation method  
480 to assess the hydro-environmental impacts of GI techniques in small urban catchments. Song  
481 and Chung (2017) utilised a daily rainfall event between 17 August 2014 and 26 August 2014 to  
482 run the hydrological model before applying multi-criteria decision analysis to prioritise site and  
483 type of GI selection. None of the studies acknowledged land-use cover change, neither past nor  
484 future.

#### 485 ***GI design and assessment***

486 Most of the publications applying social-technical perspectives (four out of five) defined  
487 GI options, including combined multiple possibilities (two out of five) or a single GI alternative  
488 (two out of five). Song and Chung (2017) tested a framework to prioritise types and sites of GI,  
489 running scenarios combining bioretention cells, rain gardens, green roofs, infiltration trench,  
490 porous pavements, and rain barrels. Yang and Chui (2018) assessed the hydro-environmental  
491 impact of bioretention cells and green roofs in different urban areas of New York.

492 Most of the publications applying a social-technical perspective (three out of five) had to  
493 simplify their assessments to deal with data scarcity and quality. Song and Chung (2017)  
494 obtained all of the design parameters of porous pavement from previous research results. Liu  
495 and Jensen (2018) collected their data from open sources (such as official websites, published  
496 plans, documents, and articles) and validated the data using an online questionnaire. O'Sullivan  
497 et al. (2015) excluded vegetation data in their model due to the lack of data for rain gardens.

498 All the studies using social-technical perspectives assessed runoff management  
499 efficiency as one of their implementation aspects. Liu and Jensen (2018) verified the GI  
500 potential for flood control and climate adaptation to reduce water footprints in Berlin and  
501 Singapore, to protect the ecosystem in Philadelphia, and to support potable water saving in  
502 Melbourne and Sino-Singapore Tianjin Eco-city. Song and Chung (2017) compared runoff  
503 reduction between infiltration trenches and porous pavements. Their results demonstrated that a  
504 single GI type can perform differently when applied to different locations and that different GI  
505 options can be more effective for different hydrological components. Cost analysis was  
506 assessed by O'Sullivan et al. (2015), who found that the incorporation of GI techniques in

507 stormwater treatment reduced the costs of running and building urban stormwater treatment  
 508 systems.

## 509 Discussion

510 To provide more evidence about how modelling approaches can produce more effective  
 511 outputs to better advise land-use planning decisions, this systematic review sought to  
 512 investigate: (i) how different approaches for modelling current and future impacts of rainfall  
 513 events incorporated GI, (ii) how models addressed differing spatial and temporal scales in their  
 514 analyses, and (iii) what aspects related to GI design and assessment were considered by  
 515 differing models. Overall, findings from this study indicate that the information generated by  
 516 current GI modelling approaches are useful and informative, but not sufficient to stretch beyond  
 517 their discipline and aid land-use planning in its policy response to climate change risks,  
 518 including floods (see Table 4 and Supplementary Material G). This gap, both between scientific  
 519 disciplines and between science and decision-makers helps explain why, despite GI's potential,  
 520 its implementation is not widespread in land-use planning.

521 *Table 4 - Key findings related to the application of exploratory questions to selected publications*

Theme	Exploratory questions	How models are currently addressing themes	
		Hydrological	Social-technical perspective
Spatial Scale	What spatial scale is covered by the study (e.g. whole catchment, single waterway)?	Most studies focused on single waterways.	Spatial scale doesn't appear to be a key consideration. Only one out of five studies assessed it, but focused on a single waterway.
	What kind of urbanisation pattern does the study focus on (e.g. urban infill, peri-urban, or greenfield development)?	Although urban infill developments are the most assessed, urbanisation patterns do not seem to be a key determinant in the choice of models used for analysis.	Although it is not a key determinant for social-technical perspectives publications, two out of five studies assessed urban infill developments as the key urbanisation pattern.
	What types of land-use are included in the study (e.g. residential, commercial, industrial)?	The majority of studies used mixed land-use types, residential being the most recurrent.	One out of five studies used land-use data.
Temporal scale	What range of historical rainfall data is used by the study?	Most of the studies used historical data and a one-decade rainfall range.	Fewer studies assessed temporal scales regarding historical rainfall data. Two out of five studies used historical rainfall data.
	What range of future rainfall data projection is used by the study (e.g. climate change projections)?	Few studies incorporated future rainfall data (seven out of 36) and among those three or more decades of rainfall data were the most used.	No climate change projections relating to future rainfall data were used by the publications.

Theme	Exploratory questions	How models are currently addressing themes	
		Hydrological	Social-technical perspective
GI design and assessment	How does the study deal with past and future land-use/land cover change?	Fewer publications assessed land-use/land-cover information (12 out of 45).	No study acknowledged land-use cover change (neither past nor future).
	How does the study design and assess GI alternatives (e.g. single option, multiple options not combined, or combined multiple options)?	Most studies tested three or more GI alternatives, bio-retention cells and permeable pavement being the most frequent.	Two out of five used combined alternatives, while the same number (two out of five) assessed single options).
	How does the study deal with data paucity and quality?	Different levels of simplification were used to calibrate and run the models.	The majority of publications simplified their assessments to deal with data paucity and quality.
	What GI implementation aspects are discussed (e.g. cost analysis, an optimal combination of GI, runoff management, suitable locations)?	The majority of studies assessed runoff management efficiency as a GI implementation aspect.	All studies focused on runoff management efficiency.

522

523 Turning to the discussion, from a spatial dimension we can see that the majority of the  
524 selected studies investigated GI options applied to single waterways (e.g. Hu et al., 2019;  
525 Schubert et al., 2017; Zhu and Chen, 2017). This is an important finding. Waterways are  
526 interconnected within a catchment and the wider urban system; therefore, to improve the quality  
527 of information for decision-making we recommend that the spatiality of studies needs to better  
528 reflect the spatiality of decision-making (Jayawardena and Marjorie, 2017). More tellingly for  
529 land-use planning, urbanisation patterns do not seem to be a key determinant utilised by the  
530 publications in the selection of models to test GI alternatives; rather their selection focused on  
531 models' capabilities and richness of functions instead. A potential solution to address these  
532 shortcomings would be the combined use of the hydrological approach with other methods,  
533 such as Technique for Order Preference by Similarity to an Ideal Solution - TOPSIS, Land  
534 Transformation Model (LTM), GIS, or others. This potentially could result in a more  
535 comprehensive assessment of the effects of extreme rainfall events at the catchment scale and  
536 guide land-use planning decisions to both deal with existing built up areas at risk and inform  
537 future developments in such areas.

538 While the advances in hydrological models does allow a straightforward analysis of the  
539 spatial distribution of floods, GI, and their effects, which has potential to inform land-use plans  
540 and policies about the spatial distribution of risks, the impacts of the implementation of GI and  
541 even the spatial understanding of the effects of rainfall events primarily focus on a single

542 catchment. This is problematic because even site-based land-use planning decisions are  
543 influenced by wider social and political dimensions and routinely address broader spatial and  
544 temporal scales, including climate change (Herath and Wijesekera, 2019; Lemes de Oliveira,  
545 2019; Ran and Nedovic-Budic, 2016). A possible way to close this gap is the co-production of  
546 tools with end-users (practitioners such as planners and flood managers), which can not only  
547 better evaluate GI techniques at larger spatial scales, but also consider more aspirational future  
548 urbanisation trends and acknowledge key political trade-offs (such as between types of land-  
549 use) in order to provide more evidence to support where GI application can minimise impacts  
550 from those land-use changes.

551 From a temporal dimension, a significant amount of publications continued to primarily  
552 focus on the historical data for both rainfall and land-use and land cover change (Chen et al.,  
553 2017; Zhang et al., 2019). While some studies developed future scenarios based on detected  
554 trends (e.g. Li et al., 2019), future changes to land-use and land cover systems were not a  
555 predominant trend in the publications analysed. Surprisingly, these aspects were even less  
556 considered by publications applying social-technical perspectives, which in theory should enable  
557 more qualitative assessments with a future thinking perspective in mind, and co-production of  
558 tools with practitioners (Jorgenson et al., 2019).

559 Future risks and uncertainty as a result of climate change are likely to increase in the  
560 future, demanding the implementation of urban adaptation measures through land-use planning  
561 (Ran and Nedovic-Budic, 2016). Higher rainfall intensities and peak flows, along with prolonged  
562 duration of rainfall events and more periodic flooding, are predicted to affect the efficiency of  
563 traditional stormwater infrastructure and have considerable implications for land-use planning as  
564 well as city budgets (Huang et al., 2018; Zevenbergen et al., 2008; Zhang et al., 2019). For  
565 example, with an increase in extreme rainfall events or a predicted increase in urbanisation,  
566 areas that do not pose a flood risk at present may be flooded in future. Notably, all studies  
567 showed limited incorporation of climate change projections, despite the potential of GI to make a  
568 significant strategic contribution in reconciling future development demands with changes in  
569 future rainfall (Carter et al., 2018; European Commission, 2013). This is also surprising as  
570 models such as SWMM allows the construction and assessment of a wide range of scenarios  
571 for simulation and analysis – a significant feature that can incorporate climate change  
572 projections. While this presents a considerable gap in knowledge for the scientific community to  
573 address, it appears that the tools needed for this (such as SWMM) are already available, but not  
574 yet used to their full potential.

575           The evaluation of the effectiveness of individual GI techniques in mitigating extreme  
576 rainfall events among the selected studies was limited, including the assessment of their  
577 potential to be implemented in multiple combinations (e.g. Bai et al., 2018; Xu and Liu, 2018;  
578 Zhu and Chen, 2017). Few publications have discussed GI effectiveness (e.g. Baek et al., 2015;  
579 Chen et al., 2017) and, when doing so, did not incorporate the spatial and temporal scales that  
580 are most useful for land-use planning. Several studies have investigated the efficacy of GI  
581 (Ahiablame and Shakya, 2016; Eaton, 2018; Mao et al., 2017), but the impact of climate change  
582 on GI effectiveness remains uncertain as impacts will likely differ significantly regarding both  
583 their spatial and temporal consequences (Sohn et al., 2019; Zhang et al., 2019). This is another  
584 area that research could usefully focus on to link to land-use planning. While it may be easier to  
585 model individual GI, in reality they would be used in combination, as their spatial demands need  
586 to reflect the differing value of land-uses across the urban area. For example, certain GI options  
587 are most appropriate where space is at a premium (e.g. planter box, permeable pavement) (Lin  
588 et al., 2018), but upstream larger options (e.g. urban forestry) may be a viable solution to reduce  
589 the demands elsewhere in the catchment (Webber et al., 2019). A systemic view of space  
590 (knowing disadvantages, potentialities and inter-connections), options (understanding costs,  
591 benefits and drawbacks), and performance (combining alternatives or using single options) is  
592 needed.

593           Finally, a more interdisciplinary approach to GI modelling (discussing uncertainties in  
594 various aspects of the land-use planning process and how these uncertainties affect decisions)  
595 is needed to support decision-making (Marot et al., 2015; Pauleit et al., 2017). Different models  
596 and approaches have sought to recognise problems in decision-making, but there is still lack of  
597 integration (Chen et al., 2017), particularly as decision-making regimes vary significantly (Scott,  
598 2019). Water plays a key role in cities and provides multiple links to other urban management  
599 areas, including land-use planning (Jayawardena, 2018). In particular, to effectively make our  
600 cities more resilient to flooding, land-use planning strategies need be based on the catchment  
601 scale, and different urban management sectors (e.g. land-use planning, stormwater  
602 management, urban infrastructure) should work more collaboratively (Hughes and Sharman,  
603 2015). For example, the use of catchment as the land-use planning scale has advantages such  
604 as self-regulation capabilities, flood control strategies and ecosystem management which can  
605 enhance flood resilience (Jayawardena, 2018). Such collaborative, holistic approaches does  
606 require improved spatial and temporal analysis of GI effectiveness, as well as information  
607 regarding GI implementation, design, and cost. Should we invest now, or wait? Discrepancy in

608 outcomes around GI implementation and uncertainty regarding climate change impacts (Gill et  
609 al., 2007) indicate that there is a key knowledge gap in the identification of GI functions and the  
610 timing and prioritisation of GI in policies, strategies and decision-making processes. There is still  
611 a degree of siloization between academic fields as well as between science and practice, which  
612 makes it difficult to widely incorporate GI (Mell and Lemes de Oliveira, 2019; O'Donnell et al.,  
613 2017). In reality, land-use planning decisions require high-resolution information at both  
614 temporal and spatial scales (Johan et al., 2004; Zhang et al., 2019). They also require thorough  
615 understanding of cross-scale processes and interactions at larger spatial scales (such as  
616 catchment scales) and across temporal scales (such as the ones related to climate change  
617 projections) (Chen et al., 2017; Laforteza et al., 2013; Yiannakou and Salata, 2017). This type  
618 of information is vital for guiding plan-making and plan-implementation, as it is more easily  
619 converted into overlays and risk assessments used to inform current and future development  
620 control and land-use planning.

621 In summary, there is no doubt the current suite of models is useful in understanding how  
622 the incorporation of GI can affect the risk of flooding. By considering these models in the context  
623 of land-use planning, this article is designed to help progress this agenda further and help  
624 sketch out a possible interdisciplinary research agenda. Decision-making concerning the use of  
625 land, particularly the need to retrofit resilience into existing urban landscapes, is a political  
626 process (Busscher et al., 2019). Evidence and risks are carefully weighted, both over different  
627 timescales and spatial scales, and over alternative land uses that can also achieve  
628 environmental, social or economic goals. Intensification may occur in one place, while  
629 greenspaces are implemented elsewhere. By better linking the two, however, we can help hit  
630 multiple goals - from climate change adaptation to amenity values for local communities. The  
631 evidence here suggests that this will involve designing research differently to better integrate  
632 decision-makers with projects, using multiple complementary models, and considering from the  
633 onset how to more deeply connect the scope of studies with the realities and trade-offs  
634 regarding current and future land-use.

## 635 **Conclusion**

636 There has been much discussion in the social science GI literature regarding  
637 implementation aspects, including its performance and maintenance (Benedict et al., 2006),  
638 reconciling the space it occupies with increasing development demands (Young, 2011),  
639 quantifying the amenity (McDonald et al., 2005), health or biodiversity values (Kambites and

640 Owen, 2006), or moving towards strategic multifunctional land-uses (Hansen and Pauleit, 2014).  
641 This study extends this discussion by linking the types of modelling approaches used and the  
642 approaches they take, with their potential to inform land-use planning, particularly in the context  
643 of precipitation extremes.

644 Adopting a systematic literature review of 50 publications, the paper assessed different  
645 approaches to modelling current and future impacts of extreme rainfall events on the urban  
646 environment, including under climatic change conditions. It found that there were two main  
647 families of modelling approaches: hydrological and social-technical perspectives, and the aim of  
648 the study and choice of modelling approaches led to very different ways to evaluate current and  
649 future impacts of extreme rainfall events. After reviewing the application of these two families of  
650 modelling approaches with respect to how they spatially and temporally incorporated GI, this  
651 study identified two key findings that can provide the basis of a future research agenda and  
652 potentially overcome barriers to GI implementation. First, the information generated by current  
653 GI modelling approaches are not well integrated into the demands of land-use planning, and  
654 may more reflect the information availability than useability. Second, there are particular gaps  
655 with regard to understanding future risks and pressures, most notably increased rainfall  
656 intensity, escalating development demands, and the interaction between the two. It was  
657 surprising to note that climate change projections and land-use futures were only rarely included  
658 in the models, despite their potentially high value for decision-makers struggling to reconcile  
659 competing demands influencing urban development.

660 Notably, the observed limitation in the inclusion of temporal and spatial scales in the  
661 analyses presents a key challenge for both flood risk management and land-use planning,  
662 especially considering increased climate change impacts. If current information does not assess  
663 the efficiency of GI in minimising future flood risks as a result of extreme rainfall events this may  
664 also limit the uptake of GI alternatives by both land-use planning and flood management  
665 sectors. In order to be better inform land-use planning decisions, hydrological models assessing  
666 GI efficiency need to incorporate a number of features. These include the incorporation of a  
667 range of temporal scales, from historical data to current and future, and the spatial scales that  
668 are most useful for planners, which are not typically single waterways but focus on changes in  
669 land-use patterns, land cover, and future development options. Decision-makers need to know,  
670 for example, which GI alternatives (or combination of alternatives) fit better in which part of the  
671 city, what are their performance under current rainfall events and climate change forecasts, how  
672 land-use interferes in hydrological parameters and what are the best allocation of GI



673 alternatives. Socio-technical approaches could also incorporate stronger social dimensions to  
674 their analysis, particularly relating to land-use constraints and climate change projections,  
675 especially because low socioeconomic areas tend to be hit the hardest by flood events. The  
676 more holistic the assessment of GI information is, the more useful and robust decisions about  
677 their widespread implementation will be.

## References

Ahiablame, L., Shakya, R., 2016. Modeling flood reduction effects of low impact development at a watershed scale. *Journal of Environmental Management* 171, 81-91.

Alves, A., Gersonius, B., Kapelan, Z., Vojinovic, Z., Sanchez, A., 2019. Assessing the Co-Benefits of green-blue-grey infrastructure for sustainable urban flood risk management. *Journal of Environmental Management* 239, 244-254.

Arnone, E., Pumo, D., Francipane, A., La Loggia, G., Noto, L.V., 2018. The role of urban growth, climate change, and their interplay in altering runoff extremes. *Hydrol Process* 32(12), 1755-1770.

Baek, S.S., Choi, D.H., Jung, J.W., Lee, H.J., Lee, H., Yoon, K.S., Cho, K.H., 2015. Optimizing low impact development (LID) for stormwater runoff treatment in urban area, Korea: Experimental and modeling approach. *Water Research* 86, 122-131.

Bai, Y., Zhao, N., Zhang, R., Zeng, X., 2018. Storm Water Management of Low Impact Development in Urban Areas Based on SWMM. *Water-Sui* 11(1).

Benedict, M.A., McMahon, E.T., Mark, A.t.C.F., Bergen, L., 2006. *Green infrastructure : Linking landscapes and communities*. Washington: Island Press, Washington.

Bilotta, G.S., Milner, A.M., Boyd, I., 2014. On the use of systematic reviews to inform environmental policies. *Environ. Sci. Policy* 42, 67-77.

Block, P., Strzepek, K., 2010. Economic Analysis of Large-Scale Upstream River Basin Development on the Blue Nile in Ethiopia Considering Transient Conditions, Climate Variability, and Climate Change. *J Water Res Plan Man* 136(2), 156-166.

Burszta-Adamiak, E., Mrowiec, M., 2013. Modelling of green roofs' hydrologic performance using EPA's SWMM. *Water Science and Technology* 68(1), 36-42.

Busscher, T., van den Brink, M., Verweij, S., 2019. Strategies for integrating water management and spatial planning: Organising for spatial quality in the Dutch "Room for the River" program. *J Flood Risk Manag* 12(1), <xocs:firstpage xmlns:xocs=""/>.

Byrne, D., 2017. What is content analysis. *Project Planner* 10, 9781526408570.

Campbell, H., 2006. Is the Issue of Climate Change too Big for Spatial Planning? *Plan Theory Pract* 7(2), 201-230.

Carter, J.G., Handley, J., Butlin, T., Gill, S., 2018. Adapting cities to climate change – exploring the flood risk management role of green infrastructure landscapes. *J Environ Plann Man* 61(9), 1535-1552.

Chen, P.Y., Tung, C.P., Li, Y.H., 2017. Low impact development planning and adaptation decision-making under climate change for a community against pluvial flooding. *Water-Sui* 9(10).

Cipolla, S.S., Maglionico, M., Stojkov, I., 2016. A long-term hydrological modelling of an extensive green roof by means of SWMM. *Ecol Eng* 95, 876-887.

Damodaram, C., Zechman, E.M., 2013. Simulation-optimization approach to design low impact development for managing peak flow alterations in urbanizing watersheds. *J Water Res Plan Man* 139(3), 290-298.

Dietrich, O., Schweigert, S., Steidl, J., Lischeid, G., 2016. Effects of Data and Model Simplification on the Results of a Wetland Water Resource Management Model. *Water-Sui* 8, 252.

Duinker, P.N., Greig, L.A., 2007. Scenario analysis in environmental impact assessment: Improving explorations of the future. *Environmental Impact Assessment Review* 27(3), 206-219.

Eaton, T.T., 2018. Approach and case-study of green infrastructure screening analysis for urban stormwater control. *Journal of Environmental Management* 209, 495-504.

Eckart, K., McPhee, Z., Bolisetti, T., 2018. Multiobjective optimization of low impact development stormwater controls. *J Hydrol* 562, 564-576.

Ercolani, G., Chiaradia, E.A., Gandolfi, C., Castelli, F., Masseroni, D., 2018. Evaluating performances of green roofs for stormwater runoff mitigation in a high flood risk urban catchment. *J Hydrol* 566, 830-845.

European Commission, 2013. Building a green infrastructure for Europe. European Commission, Belgium.

Fletcher, T.D., Shuster, W., Hunt, W.F., Ashley, R., Butler, D., Arthur, S., Trowsdale, S., Barraud, S., Semadeni-Davies, A., Bertrand-Krajewski, J.-L., Mikkelsen, P.S., Rivard, G., Uhl, M., Dagenais, D., Viklander, M., 2015. SUDS, LID, BMPs, WSUD and more – The evolution and application of terminology surrounding urban drainage. *Urban Water Journal* 12(7), 525-542.

Fu, J.C., Jang, J.H., Huang, C.M., Lin, W.Y., Yeh, C.C., 2018. Cross-Analysis of Land and Runoff Variations in Response to Urbanization on Basin, Watershed, and City Scales with/without Green Infrastructures. *Water-Sui* 10(2).

Gill, S., Handley, J.F., Ennos, R., Pauleit, S., 2007. Adapting cities for climate change: The role of the green infrastructure. *Built Environment* 33, 115-133.

Goncalves, M.L.R., Zischg, J., Rau, S., Sitzmann, M., Rauch, W., Kleidorfer, M., 2018. Modeling the effects of introducing low impact development in a tropical city: A case study from Joinville, Brazil. *Sustainability* 10(3).

Hansen, R., Pauleit, S., 2014. From multifunctionality to multiple ecosystem services? A conceptual framework for multifunctionality in green infrastructure planning for urban areas. *AMBIO* 43(4), 516-529.

Herath, H.M.M., Wijesekera, N.T.S., 2019. A state-of-the-art review of flood risk assessment in urban area, *IOP Conference Series: Earth and Environmental Science*, 1 ed.

Hislop, M., Scott, A.J., Corbett, A., 2019. What does good green infrastructure planning policy look like? Developing and testing a policy assessment tool within Central Scotland UK. *Plan Theory Pract* 20(5), 633-655.

Hu, M.C., Zhang, X.Q., Li, Y., Yang, H., Tanaka, K., 2019. Flood mitigation performance of low impact development technologies under different storms for retrofitting an urbanized area. *Journal of Cleaner Production* 222, 373-380.

Huang, C.L., Hsu, N.S., Liu, H.J., Huang, Y.H., 2018. Optimization of low impact development layout designs for megacity flood mitigation. *J Hydrol* 564, 542-558.

Hughes, J., Sharman, B., 2015. Flood resilient communities: A framework and case studies. Asia Pacific Stormwater Conference, 23.

Jayawardena, H.M.I., 2018. A research approach to develop an ecologically inspired water sensitive planning and design framework for delivering climate resilient cities: A case study of Colombo, Sri Lanka. University of Auckland.

Jayawardena, H.M.I.D.P., Marjorie, v.R., 2017. Water sensitive planning and design as an ecologically inspired approach to delivering flood resilient urban environment in Sri Lanka. *Water Practice and Technology* 12(4), 964-977.

Jia, H., Yao, H., Tang, Y., Yu, S.L., Field, R., Tafuri, A.N., 2015. LID-BMPs planning for urban runoff control and the case study in China. *Journal of Environmental Management* 149, 65-76.

Johan, A., Sten, B., Bengt, C., Graham, L.P., Göran, L., 2004. Hydrological Change – Climate Change Impact Simulations for Sweden. *Ambio* 33(4), 228-234.

Jorgenson, A.K., Fiske, S., Hubacek, K., Li, J., McGovern, T., Rick, T., Schor, J.B., Solecki, W., York, R., Zycherman, A., 2019. Social science perspectives on drivers of and responses to global climate change. *WIREs Climate Change* 10(1), e554.

Kambites, C., Owen, S., 2006. Renewed prospects for green infrastructure planning in the UK. *Planning Practice & Research* 21(4), 483-496.

Kelly, R., Sirr, L., Ratcliffe, J., 2004. Futures thinking to achieve sustainable development at local level in Ireland. *Foresight*.

Khalil, H.A.E.E., 2012. Enhancing quality of life through strategic urban planning. *Sustainable cities and society* 5, 77-86.

Kong, F.H., Ban, Y.L., Yin, H.W., James, P., Dronova, I., 2017. Modeling stormwater management at the city district level in response to changes in land use and low impact development. *Environmental Modelling & Software* 95, 132-142.

Lafortezza, R., Davies, C., Sanesi, G., Konijnendijk, C.C., 2013. Green Infrastructure as a tool to support spatial planning in European urban regions. *Iforest* 6, 102-108.

- Lamond, J., Everett, G., 2019. Sustainable Blue-Green Infrastructure: A social practice approach to understanding community preferences and stewardship. *Landscape Urban Plan* 191, 103639.
- LeGates, R., Tate, N.J., Kingston, R., 2009. Spatial thinking and scientific urban planning. *Environment and Planning B: Planning and Design* 36(5), 763-768.
- Lemes de Oliveira, F., 2019. Towards a spatial planning framework for the re-naturing of cities, in: Newman, P., Desha, C., Lemes de Oliveira, F., Mell, I. (Eds.), *Planning cities with nature*. Springer International Publishing, Cham, Switzerland, pp. 81-95.
- Li, C., Liu, M., Hu, Y., Han, R., Shi, T., Qu, X., Wu, Y., 2018. Evaluating the Hydrologic Performance of Low Impact Development Scenarios in a Micro Urban Catchment. *International Journal of Environmental Research and Public Health* 15(2).
- Li, F., Liu, Y., Engel, B.A., Chen, J., Sun, H., 2019. Green infrastructure practices simulation of the impacts of land use on surface runoff: Case study in Ecorse River watershed, Michigan. *Journal of Environmental Management* 233, 603-611.
- Liao, Z.L., He, Y., Huang, F., Wang, S., Li, H.Z., 2013. Analysis on LID for highly urbanized areas' waterlogging control: demonstrated on the example of Caohejing in Shanghai. *Water Science & Technology* 68(12), 2559-2567.
- Lin, J.Y., Chen, C.F., Ho, C.C., 2018. Evaluating the effectiveness of green roads for runoff control. *J Sustain Water Buil* 4(2).
- Liu, L., Jensen, M.B., 2018. Green infrastructure for sustainable urban water management: Practices of five forerunner cities. *Cities* 74, 126-133.
- Mallett, R., Hagen-Zanker, J., Slater, R., Duvendack, M., 2012. The benefits and challenges of using systematic reviews in international development research. *Journal of Development Effectiveness* 4(3), 445-455.
- Mao, X.H., Jia, H.F., Yu, S.L., 2017. Assessing the ecological benefits of aggregate LID-BMPs through modelling. *Ecol Model* 353, 139-149.
- Marot, N., Golobič, M., Müller, B., 2015. Green Infrastructure in Central, Eastern, and South-Eastern Europe: Is there a universal solution to environmental and spatial challenges? *Urbaniziv* 26(supplement), S1-S12.

McDonald, L., Allen, W., Benedict, M., O'connor, K., 2005. Green infrastructure plan evaluation frameworks. *Journal of Conservation Planning* 1(1), 12-43.

Mell, I., Lemes de Oliveira, F., 2019. Re-naturing our future cities, in: Newman, P., Desha, C., Lemes de Oliveira, F., Mell, I. (Eds.), *Planning cities with nature*. Springer International Publishing, Cham, Switzerland, pp. 281-285.

Moher, D., Liberati, A., Tetzlaff, J., Altman, D., 2009. Reprint-Preferred Reporting Items for Systematic Reviews and Meta-Analyses: The PRISMA Statement. *Physical Therapy* 89(9), 873-880.

Neuendorf, K.A., 2017. *The Content Analysis Guidebook, Second ed.*, Thousand Oaks, California.

O' Brien, G., O' Keefe, P., 2013. *Managing Adaptation to Climate Risk: Beyond Fragmented Responses*. London: Routledge, London.

O'Donnell, E.C., Lamond, J.E., Thorne, C.R., 2017. Recognising barriers to implementation of Blue-Green Infrastructure: a Newcastle case study. *Urban Water Journal* 14(9), 964-971.

O'Sullivan, A.D., Wicke, D., Hengen, T.J., Sieverding, H.L., Stone, J.J., 2015. Life Cycle Assessment modelling of stormwater treatment systems. *Journal of Environmental Management* 149, 236-244.

O'Donnell, E.C., Lamond, J.E., Thorne, C.R., 2017. Recognising barriers to implementation of Blue-Green Infrastructure: a Newcastle case study. *Urban Water Journal* 14(9), 964-971.

Palla, A., Gnecco, I., 2015. Hydrologic modeling of Low Impact Development systems at the urban catchment scale. *J Hydrol* 528, 361-368.

Pappalardo, V., La Rosa, D., Campisano, A., La Greca, P., 2017. The potential of green infrastructure application in urban runoff control for land use planning: A preliminary evaluation from a southern Italy case study. *Ecosyst Serv* 26, 345-354.

Paré, G., Trudel, M.-C., Jaana, M., Kitsiou, S., 2015. Synthesizing information systems knowledge: A typology of literature reviews. *Information & Management* 52(2), 183-199.

- Parra, V., Fuentes-Aguilera, P., Muñoz, E., 2018. Identifying advantages and drawbacks of two hydrological models based on a sensitivity analysis: a study in two Chilean watersheds. *Hydrological Sciences Journal* 63(12), 1831-1843.
- Pauleit, S., Zölch, T., Hansen, R., Randrup, T.B., van den Bosch, C.K., 2017. Nature-based solutions and climate change—four shades of green, in: Kabisch, N., Korn, H., Stadler, J., Bonn, A. (Eds.), *Nature-based solutions to climate change adaptation in urban areas*. Springer International Publishing, Cham, Switzerland, pp. 29-49.
- Payne, G., Payne, J., 2004. *Key Concepts in Social Research*. SAGE Publications, Ltd, London.
- Pellicer-Martínez, F., González-Soto, I., Martínez-Paz, J.M., 2015. Analysis of incorporating groundwater exchanges in hydrological models. *Hydrol Process* 29(19), 4361-4366.
- Qin, H.P., Li, Z.X., Fu, G., 2013. The effects of low impact development on urban flooding under different rainfall characteristics. *Journal of Environmental Management* 129, 577-585.
- Quay, R., 2010. Anticipatory Governance. *J Am Plann Assoc* 76(4), 496-511.
- Ran, J., Nedovic-Budic, Z., 2016. Integrating spatial planning and flood risk management: A new conceptual framework for the spatially integrated policy infrastructure. *Comput Environ Urban* 57, 68-79.
- Rusli, S.R., Yudianto, D., Liu, J.-t., 2015. Effects of temporal variability on HBV model calibration. *Water Sci Eng* 8(4), 291-300.
- Salata, K., Yiannakou, A., 2016. Green infrastructure and climate change adaptation. *Tema* 9(1), 10-27.
- Schifman, L.A., Herrmann, D.L., Shuster, W.D., Ossola, A., Garmestani, A., Hopton, M.E., 2017. Situating green infrastructure in context: A framework for adaptive socio-hydrology in cities. *Water Resour Res* 53(12), 10139-10154.
- Schubert, J.E., Burns, M.J., Fletcher, T.D., Sanders, B.F., 2017. A framework for the case-specific assessment of Green Infrastructure in mitigating urban flood hazards. *Adv Water Resour* 108, 55-68.

Schuch, G., Serrao-Neumann, S., Morgan, E., Low Choy, D., 2017. Water in the city: Green open spaces, land use planning and flood management – An Australian case study. *Land Use Policy* 63, 539-550.

Scott, A., 2019. Mainstreaming the environment in planning policy and decision making, in: Davoudi, S., Cowell, R., White, I., Blanco, H. (Eds.), *The routledge companion to environmental planning*. Milton: Taylor & Francis Group, pp. 420-433.

Sohn, W., Kim, J.H., Li, M.H., Brown, R., 2019. The influence of climate on the effectiveness of low impact development: A systematic review. *Journal of Environmental Management* 236, 365-379.

Song, J.Y., Chung, E.S., 2017. A Multi-Criteria Decision Analysis System for Prioritizing Sites and Types of Low Impact Development Practices: Case of Korea. *Water-Sui* 9(4).

Soz, S.A., Kryspin-Watson, J., Stanton-Geddes, Z., 2016. The role of green infrastructure solutions in urban flood risk management. *Urban Floods Community of Practice*.

Sun, Y.W., Li, Q.Y., Liu, L., Xu, C.D., Liu, Z.P., 2014. Hydrological simulation approaches for BMPs and LID practices in highly urbanized area and development of hydrological performance indicator system. *Water Sci Eng* 7(2), 143-154.

Thorne, C.R., Lawson, E.C., Ozawa, C., Hamlin, S.L., Smith, L.A., 2018. Overcoming uncertainty and barriers to adoption of Blue-Green Infrastructure for urban flood risk management. *J Flood Risk Manag* 11(S2), S960-S972.

Wang, M., Zhang, D., Cheng, Y., Tan, S.K., 2019. Assessing performance of porous pavements and bioretention cells for stormwater management in response to probable climatic changes. *Journal of Environmental Management* 243, 157-167.

Wang, M., Zhang, D.Q., Adhityan, A., Ng, W.J., Dong, J.W., Tan, S.K., 2016. Assessing cost-effectiveness of bioretention on stormwater in response to climate change and urbanization for future scenarios. *J Hydrol* 543, 423-432.

Ward, S., Staddon, C., De Vito, L., Zuniga-Teran, A., Gerlak, A.K., Schoeman, Y., Hart, A., Booth, G., 2019. Embedding social inclusiveness and appropriateness in engineering assessment of green infrastructure to enhance urban resilience. *Urban Water Journal* 16(1), 56-67.



- Webber, J.L., Fletcher, T.D., Cunningham, L., Fu, G., Butler, D., Burns, M.J., 2019. Is green infrastructure a viable strategy for managing urban surface water flooding? *Urban Water Journal*, 1-11.
- White, I., Haughton, G., 2017. Risky times: Hazard management and the tyranny of the present. *International Journal of Disaster Risk Reduction* 22, 412-419.
- White, I., Nandedkar, G., 2019. The housing crisis as an ideological artefact: Analysing how political discourse defines, diagnoses, and responds. *Housing Studies*, 1-22.
- Woodward, M., Kapelan, Z., Gouldby, B., 2014. Adaptive flood risk management under climate change uncertainty using real options and optimization. *Risk Analysis* 34(1), 75-92.
- Xu, D., Liu, Y., 2018. Research on the effect of rainfall flood regulation and control of wetland park based on SWMM model - A case study of wetland park in Yuanjia village, Qishan county, Shaanxi province, *IOP Conference Series: Earth and Environmental Science*, 5 ed.
- Yang, Y., Chui, T.F.M., 2018. Integrated hydro-environmental impact assessment and alternative selection of low impact development practices in small urban catchments. *Journal of Environmental Management* 223, 324-337.
- Yau, W.K., Radhakrishnan, M., Liong, S.Y., Zevenbergen, C., Pathirana, A., 2017. Effectiveness of ABC waters design features for runoff quantity control in Urban Singapore. *Water (Switzerland)* 9(8), 577.
- Yiannakou, A., Salata, K.-D., 2017. Adaptation to climate change through spatial planning in compact urban areas: a case study in the city of Thessaloniki. *Sustainability* 9(2), 271.
- Young, R.F., 2011. Planting the living city. *J Am Plann Assoc* 77(4), 368-381.
- Zevenbergen, C., Veerbeek, W., Gersonius, B., van Herk, S., 2008. Challenges in urban flood management: travelling across spatial and temporal scales. *J Flood Risk Manag* 1(2), 81-88.
- Zhang, K.F., Manuelpillai, D., Raut, B., Deletic, A., Bach, P.M., 2019. Evaluating the reliability of stormwater treatment systems under various future climate conditions. *J Hydrol* 568, 57-66.

Zhang, N., Luo, Y.J., Chen, X.Y., Li, Q., Jing, Y.C., Wang, X., Feng, C.H., 2018. Understanding the effects of composition and configuration of land covers on surface runoff in a highly urbanized area. *Ecol Eng* 125, 11-25.

Zhu, Z.H., Chen, X.H., 2017. Evaluating the effects of low impact development practices on urban flooding under different rainfall intensities. *Water-Sui* 9(7).